A NOVEL ROUTE FOR WASTE WATER TREATMENT: PHOTO-ASSISTED FENTON DEGRADATION OF NAPHTHOL GREEN B

Anil KUMAR^{*a*1}, Mukesh PALIWAL^{*a*2}, Rameshwar AMETA^{*b*1} and Suresh C. AMETA^{*b*2,*}

 ^a Photochemistry and Solar Energy Laboratory, Department of Chemistry, University College of Science, M. L. Sukhadia University, Udaipur 313001, India; e-mail: ¹ anilchohadia@yahoo.co.in, ² paliwalmk_2005@yahoo.co.in

^b Department of Chemistry, Government Meera Girls College, Udaipur 313001, India; e-mail: ¹ ameta_ra@yahoo.com, ² ameta_sc@yahoo.com

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This study was conducted to assess the removal of the Naphthol Green B dye from aqueous medium using the photo-Fenton process. The Fenton reagent, a mixture of hydrogen peroxide and Fe³⁺ ions, was used to generate the hydroxyl radical ('OH) that degrades the dye. Experiments were conducted at laboratory temperature and atmospheric pressure to examine the effect of reaction conditions such as the concentration of Fe³⁺ ions, the dye and hydrogen peroxide, pH, and light intensity on the reaction rate. The progress of the photochemical degradation was monitored spectrophotometrically. The optimum photochemical degradation conditions were determined. Naphthol Green B was completely degraded into CO_2 and H_2O . A tentative mechanism for photochemical bleaching of the dye by the photo-Fenton reaction has been proposed.

Keywords: Photodegradation; Photo-Fenton process; Naphthol Green B; Hydrogen peroxide.

Many dyes are toxic and carcinogenic. Contamination of the environment by toxic chemicals is emerging as a serious global problem. Coloured effluents containing dyes from textile, dyeing and printing industries may cause skin cancer due to photosensitisation and photodynamic damage. On the contrary, bleached dye solutions are non-toxic and harmless. Second, the dye-containing coloured water is hardly of any use, but if the coloured solutions are bleached to give colourless water, it may be used for washing, cooling, irrigation and cleaning purposes. Thus, photochemical bleaching provides a low-cost method to solve the problem of water pollution.

The possibility of using Fenton-based systems in dark reactions^{1,2} and under light irradiation³⁻⁶ has been a topic of recent interest in the field of advanced oxidation technologies involving degradation of pollutants

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in homogeneous systems. The Fenton reactions in the dark mineralise ca. 40–60% of the total organic carbon (TOC) present^{7–10}. Recently, it has been found that illumination of the Fe^{2+} – Fe^{3+} – H_2O_2 system increases the degradation rate of many aromatic organic substances, such as phenols, substituted phenols, nitrobenzene, anisole^{11–14}, TNT and RDX^{15,16}, Orange II and 4-chloroaniline¹⁷.

The Fenton reaction 18 produces hydroxyl radicals by the reaction of $\rm H_2O_2$ with $\rm Fe^{2+}$ salts.

$$Fe^{2+} + H_2O_2 \longrightarrow Fe^{3+} + OH + OH^-$$
 (1)

In the dark, the reaction is retarded after complete conversion of Fe^{2+} to Fe^{3+} . The positive effect of illumination on the rate of degradation is due to the photoreduction of Fe^{3+} to Fe^{2+} ions producing new 'OH radicals (and H_2O_2) according to Eq. (2).

$$Fe^{3+} + H_2O \xrightarrow{hv} Fe^{2+} + OH + H^+$$
 (2)

The direct photolysis of H₂O₂ also generates 'OH radicals.

$$H_2O_2 \xrightarrow{hv} 2 OH \qquad (\lambda < 400 nm)$$
 (3)

Chen et al.¹⁹ studied the photochemical degradation of azo dyes, Red MX-5B, Reactive Black-5 and Orange G, using low iron concentrations in the Fenton and Fenton-like systems. Daneshvar and Khataee²⁰ decolourised solution containing a common textile and leather dye, C.I. Acid Red 14 (AR14), at pH 3, using the Fenton, $UV/H_2O_2/O_2$, $UV/H_2O_2/Fe^{2+}$, $UV/H_2O_2/Fe^{3+}$ and UV/H₂O₂/Fe³⁺/oxalate processes. The influence of Alizarin Violet 3B on the Fenton reaction of organic compounds under irradiation ($\lambda > 450$ nm) was examined by Tang et al.²¹. Chaudhuri et al.²² also chemically oxidised Methylene Blue using a Fenton-like reaction. Prousek et al.²³ reported the utilisation of the Fenton reaction for degradation of dyes present in coloured waste water. Photodegradation of Malachite Green in the presence of Fe^{3+}/H_2O_2 under irradiation with visible light was observed by Wu et al.²⁴. Ruppert et al.²⁵ studied the photo-Fenton reaction as an effective photochemical process to treat polluted water. Mogra et al.^{26,27} reported the photochemical degradation of 1,4-dichlorobenzene and chlorobenzene by the photo-Fenton reagent. Walling²⁸ studied intermediates in the reaction of Fenton-type reagents.

The Fenton reaction is easy and forms no sludge. It has been therefore widely used to degrade pollutants^{29,30}. Hydroxyl radical is a powerful oxi-

dant which can rapidly and non-selectively oxidise organic contaminants to carbon dioxide and water^{31,32}; hence, it is able to degrade pollutants effectively^{33,34}. The photo-Fenton reaction is another effective method of preventing pollution. Xiang and Zheng³⁵ accelerated decolourisation of dyes by combination of UV radiation and Fenton process, which produced the 'OH radicals directly (Eqs (4)–(6)).

$$H_2O_2 \xrightarrow{hv} 2 OH$$
 (4)

$$OH + dye \longrightarrow dye intermediate$$
 (5)

•OH + dye intermediate \longrightarrow CO₂ + H₂O + mineralisation products (6)

EXPERIMENTAL

Naphthol Green B (Sisco), FeCl₃ (CDH) and H₂O₂ (30%, Merck), were used. Photochemical degradation of the dye was studied in the presence of Fe³⁺ ions, H₂O₂ and visible light. Naphthol Green B (0.0879 g) was dissolved in double distilled water (100 ml) to give 1.0 \times 10^{-4} M solution. Anhydrous FeCl₃ (0.4055 g) was dissolved in double distilled water (500 ml) to give 5.0×10^{-3} M stock solution. Photochemical degradation of Naphthol Green B was investigated in 30 ml of reaction solution that contained the dye, FeCl₃ and H₂O₂ at concentrations 3.33×10^{-6} , 1.0×10^{-6} and 1.00×10^{-3} mol/l, respectively. The reaction mixture was exposed to the light (70.0 mW/cm²), of a 200 W tungsten lamp (Sylvania Laxman). Sunlight was used for higher light intensities. The light intensity was measured with a Suryamapi CEL Model SM 201 powermeter. A water filter was used to cut off thermal radiation. pH of the solution was adjusted by the addition of standardised hydrochloric acid and sodium hydroxide solutions and measured with a Systronics Model 106 digital pH meter. The necessary condition for the correct measurement of absorbance is that the solution must be free of solids. Therefore, a Remi-1258 centrifuge and a Whatman filter paper were attempted to remove the solids but both failed. Hence, a G 3 sintered glass crucible was used for filtration to obtain the desired accuracy in recording the absorbance of the dye with a Shimadzu U.V. 240 spectrophotometer.

RESULTS AND DISCUSSION

The photochemical degradation of Naphthol Green B (Chart 1) was monitored at $\lambda_{max} = 710$ nm. The results of a typical run are given in Table I and Fig. 1.

It was observed that absorbance (*A*) of Naphthol Green B solutions decreased with irradiation time, indicating that the dye was consumed. A plot of $1 + \log A$ against time is linear, consistent with pseudo-first-order kinet-

ics. The rate constant *k* was calculated from the expression $k = 2.303 \times$ slope. The optimum rate constant for this reaction was determined as $k = 16.88 \times 10^{-4} \text{ s}^{-1}$.



Chart 1

Effect of pH

The effect of pH on the rate of degradation of Naphthol Green B was investigated in the pH range 2.4–3.8. The results are reported in Table II.

The photochemical degradation depends strongly on pH of the reaction medium as evident from the data in Table II. The reaction rate increases with pH increasing up to 3.2 and then it decreases. The hydroxyl radicals are generated in two steps:

(i) in the reaction between Fe^{2+} and hydrogen peroxide,

(ii) in the photochemical reaction of Fe^{3+} ions with water.



Fig. 1

A typical run of degradation of Naphthol Green B by the photo-Fenton process. Experimental conditions are given in Table I. The slope corresponds to $k = 16.88 \times 10^{-4} \text{ s}^{-1}$

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The increase in pH of the medium favours the step (i) where OH⁻ ions are formed along with hydroxyl radicals, whereas protons are generated in the step (ii). Thus, it may be concluded that the step (i) dominates over the step (ii) at pH below 3.2, and vice versa at pH above 3.2.

<i>t</i> , min	A	$1 + \log A$
0	0.876	0.9426
1	0.822	0.9150
2	0.768	0.8855
3	0.714	0.8538
4	0.660	0.8195
5	0.606	0.7825
6	0.552	0.7422
7	0.498	0.6974
8	0.444	0.6474
9	0.390	0.5912
10	0.336	0.5266
11	0.317	0.5021
12	0.299	0.4761
13	0.271	0.4334
14	0.242	0.3854
15	0.214	0.3314
16	0.194	0.2894
17	0.167	0.2248
18	0.152	0.1833
19	0.135	0.1316
20	0.123	0.0913

TABLE I A typical run of degradation of Naphthol Green B by the photo-Fenton process^a

^a Experimental conditions: [Naphthol Green B] = 3.33×10^{-6} mol/l, [H₂O₂] = 1.00×10^{-3} mol/l, [Fe³⁺] = 6.67×10^{-4} mol/l, pH 3.2, incident light intensity = 70.0 mW/cm².

1 1 0	1
рН	$k \times 10^4$, s ⁻¹
2.4	4.30
2.6	7.39
2.8	13.16
3.0	15.76
3.2	16.89
3.4	4.93
3.6	0.50
3.8	0.32

The effect of pH on the rate of photodegradation of Naphthol Green B^a

^a Experimental conditions: [Naphthol Green B] = $3.33 \times 10^{-6} \text{ mol/l}$, $[\text{H}_2\text{O}_2] = 1.00 \times 10^{-3} \text{ mol/l}$, $[\text{Fe}^{3+}] = 6.67 \times 10^{-4} \text{ mol/l}$, incident light intensity = 70.0 mW/cm².

Effect of Naphthol Green B Concentration

Effect of varied dye concentration on the reaction rate was studied for reaction conditions given in Table III.

The rate of photochemical degradation was found to increase with increasing concentration of Naphthol Green B up to 3.33×10^{-5} mol/l. On

TABLE III			
The effect of Naphthol Green B	concentration on the	rate of photodegradation	of the dye^a

$[NBG] \times 10^6$, mol/l	$k \times 10^4$, s ⁻¹	
0.67	10.55	
1.33	12.72	
2.00	15.19	
2.67	16.18	
3.33	16.89	
4.00	10.55	
4.67	9.85	
5.33	9.85	
6.00	8.97	
6.67	6.33	

 a Experimental conditions: $[\rm H_2O_2]$ = 1.00 \times 10 $^{-3}$ mol/l, $[\rm Fe^{3+}]$ = 6.67 \times 10 $^{-4}$ mol/l, pH 3.2, incident light intensity = 70.0 mW/cm².

TABLE II

further increasing its concentration, a sudden decrease in the rate of degradation was observed. This result may be attributed to the fact that at increased concentration Naphthol Green B may act itself as a filter for the incident light. The high concentration does not permit the light to reach the dye molecules in the bulk solution.

Effect of Fe³⁺ Concentration

The effect of concentration of Fe^{3+} ions on the rate of photochemical degradation of Naphthol Green B was studied by keeping all the other factors identical. The results are summarised in Table IV.

It is clear from the experimental data that the rate of photodegradation increases with the increasing concentration of Fe³⁺ ions up to 6.67 × 10^{-5} mol/l. This is tentatively explained by increased concentration of Fe²⁺ ions in parallel with Fe³⁺, which is accompanied by enhanced generation of 'OH radicals.

TABLE IV

The effect of Fe³⁺ concentration on the rate of photodegradation of Naphthol Green B^a

$[Fe^{3+}] \times 10^4$, mol/l	$k \times 10^4$, s ⁻¹	
0.67	0.88	
1.33	1.82	
2.00	2.60	
2.67	4.16	
3.33	5.98	
4.00	8.66	
4.67	9.96	
5.33	11.43	
6.00	13.64	
6.67	16.89	

^a Experimental conditions: [Naphthol Green B] = 3.33×10^{-6} mol/l, [H₂O₂] = 1.00×10^{-3} mol/l, pH 3.2, incident light intensity = 70.0 mW/cm².

Effect of Hydrogen Peroxide

The effect of the amount of hydrogen peroxide on the photodegradation of Naphthol Green B was also investigated. The results are given in Table V.

The effect of hydrogen peroxide on the rate of photodegradation of Naphthol Green B^{a}

$[H_2O_2] \times 10^3$, mol/l	$k \times 10^4$, s ⁻¹
0.29	11.26
0.43	11.26
0.57	12.48
0.72	14.33
0.86	15.01
1.00	16.89
1.15	13.77
1.29	12.77
1.43	11.86
2.15	9.79

^a Experimental conditions: [Naphthol Green B] = 3.33×10^{-6} mol/l, [Fe³⁺] = 6.67×10^{-4} mol/l, pH 3.2, incident light intensity = 70.0 mW/cm².

It was observed that the reaction rate increased with the concentration of H_2O_2 , reaching an optimum value 1.00×10^{-3} mol/l. Thereafter, the rate of degradation decreased. This behaviour can be explained by the rates of the reactions (7) and (8). In Eq. (8), 'OH radicals are consumed rapidly by encountering move H_2O_2 molecules. From Eqs (7) and (8), 'OOH radicals are generated in significant amounts. They react with Fe³⁺ (Eq. (9)) and protons are liberated. The production of H⁺ is confirmed by a slight decrease in pH of the reaction mixture at the end of photoreaction. As a consequence, the rate of photodegradation decreases.

$$Fe^{3+} + H_2O_2 \xrightarrow{hv} Fe^{2+} + OOH + H^+$$
(7)

$$^{\circ}OH + H_2O_2 \longrightarrow ^{\circ}OOH + H_2O \tag{8}$$

$$Fe^{3+} + OOH \longrightarrow Fe^{2+} + O_2 + H^+$$
 (9)

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Effect of Light Intensity

The effect of light intensity on the photodegradation of Naphthol Green B was also investigated. The results obtained are reported in Table VI.

A plot of the rate constant versus incident light intensity is linear, which indicates that increasing light intensity accelerates the process. This behaviour may be attributed to higher concentration of active hydroxyl radicals formed according to Eq. (2).

TABLE VI

The effect of incident light intensity on the rate of photodegradation of Naphthol Green B^a

Light Intensity, mW/cm ²	$k \times 10^4$, s ⁻¹	
10.0	10.69	
20.0	12.54	
30.0	13.93	
40.0	14.88	
50.0	15.62	
60.0	16.20	
70.0	16.89	

 a Experimental conditions: [Naphthol Green B] = 3.33 \times 10⁻⁶ mol/l, [H₂O₂] = 1.00 \times 10⁻³ mol/l, [Fe³⁺] = 6.67 \times 10⁻⁴ mol/l, pH 3.2.

Mechanism

On the basis of our experimental evidence and the existing literature, a tentative mechanism has been proposed for the photodegradation of Naphthol Green B with the photo-Fenton reagent.

$$Fe^{3+} + H_2O \longrightarrow Fe^{2+} + OH + H^+$$
 (10)

$$\operatorname{Fe}^{3+} + \operatorname{H}_2\operatorname{O}_2 \xrightarrow{h_V} \operatorname{Fe}^{2+} + \operatorname{OOH} + \operatorname{H}^+$$
 (11)

$$Fe^{2+} + H_2O_2 \longrightarrow Fe^{3+} + OH + OH^-$$
(12)

$$OH + H_2O_2 \longrightarrow OOH + H_2O$$
(13)

$$Fe^{2+} + OH \longrightarrow Fe^{3+} + OH$$
(14)

$$Fe^{3+} + OOH \longrightarrow Fe^{2+} + O_2 + H^+$$
(15)

$$OH + OH \longrightarrow H_2O_2$$
 (16)

An aqueous solution of ferric ions decomposes water on exposure to light into proton and 'OH radical, and ferric ions are reduced to ferrous ions. The ferrous ions decompose H_2O_2 into hydroxide ion and hydroxyl radical and reoxidise to ferric ions. The reaction of ferric ions with H_2O_2 upon irradiation produces 'OOH radicals. The reaction of 'OH with H_2O_2 also produces 'OOH radicals. Ferrous ions will undergo oxidation to ferric ions by 'OH radicals, while ferric ions get reduced to ferrous ions by 'OOH radicals. The 'OOH radicals are highly unstable in water and undergo facile disproportionation rather than reacting slowly with the dye molecules. The participation of hydroxyl radicals as active oxidising species was confirmed by using a hydroxyl radical scavenger like propan-2-ol. As a result, the rate of photodegradation was drastically reduced.

The 'OH radicals are consumed in four different reactions. They can induce dissociation of H_2O_2 into 'OOH and water or they combine to form H_2O_2 molecules. Most importantly, they react with Naphthol Green B to give colourless degradation products.

The main advantage of using photo-Fenton reagent is the regeneration of the consumed Fe^{2+} ions on illumination. Each Fe^{2+} ion can produce many 'OH radicals in contrast to the dark Fenton reaction. The process is a cyclic one, only a single 'OH radical being formed per one ferrous ion. This means that the amount of ferrous salt required under photo-Fenton conditions is small compared to that under Fenton conditions, where ferrous ions need to be added in required intervals otherwise the reaction stops after complete conversion of ferrous ions to ferric ones. This is important for industrial application as further separation of ferric ions is not required after the waste water treatment^{36–41}.

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